

1 **Magnetic porous carbons derived from cobalt (II)-based metal-organic frameworks**
2 **for the solid-phase extraction of sulfonamides**

3

4 Sandra Yadira Mendiola Alvarez,^{1,2} Gemma Turnes Palomino,² Jorge Guzmán Mar,¹ Ma.
5 Aracely Hernández Ramírez,¹ Laura Hinojosa Reyes*,¹, Carlos Palomino Cabello*,²

6

7 ¹ *Facultad de Ciencias Químicas, Universidad Autónoma de Nuevo León, UANL, Cd.*
8 *Universitaria, C.P. 66455 San Nicolás de los Garza, Nuevo León, México*

9 ²*Department of Chemistry, University of the Balearic Islands, E-07122 Palma de Mallorca,*
10 *Spain*

11 **E-mail: carlos.palomino@uib.es *E-mail: laura.hinojosary@uanl.edu.mx*

12

13

14

15

16

17

18

19

20

21

22

23

24

25

26

27

28

29 **ABSTRACT**

30 In this work, the dispersive solid-phase extraction of sulfonamide antibiotics was evaluated
31 using magnetic porous carbon derived from cobalt (II)-based metal-organic frameworks. By
32 direct carbonization under inert atmosphere of Co-SIM-1, Co-MOF-74 and Co-DABCO
33 MOFs, different magnetic porous carbons were prepared and characterized to study their
34 structural, morphological, chemical and textural properties. Their performance for the
35 simultaneous extraction of three sulfonamides (sulfadiazine, sulfamerazine and
36 sulfamethazine), prior to HPLC analysis, was also evaluated, obtaining the best results (>
37 95%) in the case of C/Co-SIM-1 carbon, probably due to its bimodal pore structure, high
38 surface area and high amount of surface defects. Using this adsorbent, the effect of the pH
39 solution and contact time on the adsorption of the sulfonamides, as well as the reusability of
40 the carbon were studied.

41

42 **Keywords:** Porous carbons, metal-organic frameworks, sulfonamides, magnetic solid-phase
43 extraction, water pollutants.

44 **1. Introduction**

45 The consumption of antibiotics is increasing in both industrialized and developing countries,
46 resulting in their frequent detection in surface waters around the world. Sulfonamides,
47 classified as emerging contaminants (EC), are synthetic broad-spectrum antibiotics, which
48 are widely used in veterinary to inhibit both gram-positive and gram-negative bacteria, as
49 well as some protozoa, and also in animal industry as growth promoters.^{1,2} This kind of
50 antibiotics, also known as sulfa drugs, is not biodegradable and, due to their high polarity
51 and water solubility, usually cannot be completely removed from water in conventional water
52 treatment plants, having been detected in some water bodies.³⁻⁵ The long-term effects of low
53 concentrations of these pollutants, due to its biological activity and stability in water, can
54 pose serious risks to human health and to the environment.

55 Among the various available methods developed to remove organic pollutants, adsorption
56 has emerged as one of the most attractive because of its simplicity, reliability, and cost
57 effectiveness without generating secondary pollutants.^{6,7} Owing to their unique
58 characteristics, such as high surface area, variable functionality and tunable pore size and
59 composition, metal-organic frameworks (MOFs) have been applied as adsorbents of different
60 pollutants, including some sulfonamides. For instance, the extraction of sulfamethoxazole by
61 MIL-53 MOFs with different metallic centers⁸ or the excellent performance of ZIF-67 and
62 UiO-66 MOFs for the adsorptive removal of sulfachloropyradazine antibiotic in wastewater⁹
63 have been described, while metal-organic frameworks with coordinatively unsaturated sites,
64 such as HKUST-1 and MIL-101, have been studied as efficient sorbents for the adsorption
65 of different sulfonamides.^{10,11}

66 Although many MOFs have been tested on the extraction of organic pollutants, several of
67 them have been described as unstable in aqueous solutions under a wide range of pH
68 conditions, limiting their applicability as adsorbents.¹² In addition, the laborious and time-
69 consuming retrieval of these porous materials from the sample medium, inhibits its direct
70 application for the extraction of organic pollutants in batch conditions.

71 Recently, stable porous carbons have been prepared by a facile and single-step calcination
72 procedure under inert atmosphere without the need of any additional carbon precursor using
73 metal-organic frameworks as both templates and carbon precursors.^{13,14} The high porosity,
74 defined structure and outstanding designability of MOFs allow a precise control of the size,
75 structure, morphology and composition of the MOF-derived porous carbon. The obtained
76 carbons are uniform doped with heteroatoms from the organic linkers, which can act as active
77 sites, show excellent chemical and thermal stability, high porosity and hydrophobicity which
78 makes them good candidates as sorbents for the removal of organic pollutants.¹⁵⁻¹⁷
79 Furthermore, bearing in mind that after the carbonization process the metallic component of
80 the MOF aggregates, shaping metal or metal oxide nanoparticles, the use of iron, nickel, and
81 cobalt based MOFs as precursors to prepare porous carbons, results in magnetic porous
82 carbons, which facilitates their retrieval due to the easy magnetic separation of the adsorbent
83 avoiding filtration or centrifugation steps.¹⁸⁻²¹

84 Thus, the aim of this work is to explore the use of magnetic porous carbons, obtained by
85 simple one-step pyrolysis of different cobalt (II)-based metal-organic frameworks (Co-SIM-
86 1, Co-MOF-74 and Co-DABCO), for the adsorption of sulfonamides (sulfamethazine,
87 sulfamerazine and sulfadiazine). The influence of the pH of the extraction medium on the
88 adsorption mechanism of sulfonamides compounds from water was established and, taking

89 into account its practical application, kinetics, regeneration and reusability of the carbon
90 showing highest adsorption capacity (C/Co-SIM-1) were also examined. The obtained results
91 demonstrate that the C/Co-SIM-1 is a promising adsorbent for the efficient sulfonamides
92 extraction from water systems.

93 **2. Experimental section**

94 **2.1 Chemicals**

95 All chemicals used were of analytical grade. N,N-dimethylformamide (DMF, extrapure) and
96 ethanol (99%) were obtained from Scharlau (Barcelona, Spain). Cobalt(II) nitrate
97 hexahydrate (99%), 4-methyl-5-imidazolecarboxaldehyde (99%), 2,5-dihydroxyterephthalic
98 acid (98%), 1,4-diazabicyclo[2.2.2]octane (DABCO, 98%), sulfamethazine sodium salt
99 (SMTZ, 98%), sulfamerazine (SMRZ, 99%), sulfadiazine (SDZ, 99%), trifluoroacetic acid
100 (TFA, 99%) were purchased from Sigma Aldrich. Acetonitrile (HPLC grade) was obtained
101 from PanReac AppliChem. Methanol (HPLC grade) was acquired from Tedia. Milli-Q water
102 was obtained from a Direct-8 purification system (resistivity > 18 M Ω cm, Millipore Iberica,
103 Spain).

104 **2.2 Instrumentation**

105 The morphology of the prepared materials was examined by scanning electron microscopy
106 using a Hitachi S-3400N microscope, equipped with a Bruker AXS Flash 4010 energy-
107 dispersive X-ray spectroscopy (EDS) system. X-ray diffraction (XRD) patterns were
108 obtained on a Bruker D8-Advanced diffractometer using Cu-K α radiation ($\lambda = 1.5418 \text{ \AA}$)
109 from 5-80° with a step size of 0.02 (2 θ). Raman spectra were acquired from 100 to 1000 cm⁻¹
110 on a NRS-5100 (JASCO Co., Japan) using a charge coupled device (CCD) detector at the
111 laser excitation of 785 nm with 10.8 mW laser power. The I_D/I_G ratios were calculated by

112 measuring the intensity on the band maximum. The nitrogen adsorption/desorption isotherms
113 were collected at 77 K using a micromeritics TriStar II 3020 apparatus (Micromeritics
114 Instrument Corporation). The samples were degassed at 423 K for 12 h using the outgas port
115 of the TriStar II instrument. The isotherms were analyzed by using the Brunauer-Emmett-
116 Teller (BET) method to determine the specific surface area and the two dimensional non-
117 local density functional theory (2D-NLDFT) model for the determination of the pore volume
118 and the pore size distribution. The isoelectric point was measured in a Zetasizer Nano ZS90
119 (Malvern) by dispersing 5 mg of carbon samples into 10 mL of water at 25 °C and adjusting
120 the pH of the suspension with 0.1 M HCl or 0.1 M NaOH. X-ray photoelectron spectroscopy
121 (XPS) measurements were performed on a PHI 5000 Versa Probe II XPS system using a
122 monochromatic Al K α X-ray source (hv = 1486.7 eV). The analysis of sulfonamides (SNs)
123 by HPLC was carried out using an HPLC Jasco MD-4017 equipped with a UV-Vis diode
124 array detector. The separation of SNs was carried out by isocratic elution with water acidified
125 with 0.0025% of TFA-acetonitrile (90:10, v/v) on a C18 KINETEX phenomenex[®] 5 μ m (150
126 \times 4.6 mm) column at a flow rate of 0.7 mL/min. The analytes were detected at 230 nm.

127 **2.3 Synthesis of cobalt (II)-based metal-organic frameworks**

128 Co-SIM-1, Co-MOF-74 and Co-DABCO were synthesized by a solvothermal route using
129 experimental conditions previously described.²²⁻²⁴ Briefly, Co-SIM-1 was prepared by
130 dissolving 3.01 g of 4-methyl-5-imidazolecarboxaldehyde and 1.99 g of cobalt(II) nitrate
131 hexahydrate in 50 mL of DMF under constant stirring. The mixture was introduced into a
132 wide mouth glass jar and heated at 358 K for 72 h. The resulting powder was washed 3 times
133 with DMF and ethanol and then dried at 373 K overnight.

134 For preparation of Co-MOF-74, 0.75 g of 2,5-dihydroxyterephthalic acid and 3.56 g of
135 cobalt(II) nitrate hexahydrate were added successively to 300 mL of a solution of
136 DMF:ethanol:water (1:1:1, v/v) under constant magnetic stirring, which was maintained for
137 1 h. The resulting solution was transferred into a wide mouth glass jar and heated at 373 K
138 in an oven for 24 h. After cooling to room temperature, the red solid was filtered and washed
139 3 times with DMF, and then soaked in methanol for 2 days (the methanol was replaced 4
140 times during this period). Finally, the solid was filtered and then dried at room temperature
141 under vacuum.

142 Co-DABCO MOF was synthesized by mixing a cobalt(II) nitrate solution (0.34 g of
143 cobalt(II) nitrate hexahydrate in 10 mL of DMF) with 0.21 g of 2-aminoterephthalic acid
144 dissolved in 10 mL of DMF under stirring. After that, a DABCO solution (0.21 g of DABCO
145 in 10 mL of DMF) was added with constant stirring, that was maintained for 5 h. After this
146 time the solid formed was separated by filtration and the solution was heated at 393 K in a
147 stainless-steel autoclave for 72 h. The resulting powder was washed thoroughly with DMF
148 and dried at room temperature for 12 h.

149 **2.4 Preparation of magnetic carbons**

150 The magnetic porous carbons C/Co-SIM-1, C/Co-MOF-74 and C/Co-DABCO were obtained
151 by direct calcination of 2 g of the synthesized Co-SIM-1, Co-MOF-74 and Co-DABCO,
152 respectively, in a tubular furnace under a nitrogen gas flow at 1073 K (heating rate of 3 K
153 min⁻¹) for 2.5 h. The materials were then cooled to room temperature maintaining the N₂
154 atmosphere. Finally, the obtained carbons were washed with water and dried under vacuum.

155 **2.5 Solid-phase extraction procedure**

156 The extraction experiments were conducted under batch conditions using 50 mL of
157 sulfonamides (SNs) solutions containing 5 mg/L of each compound (SDZ, SMTZ and
158 SMRZ) and 5 mg of magnetic porous carbon as adsorbent. After 4 h of stirring, an external
159 magnet was placed outside of the sample beaker to separate the sorbent from the aqueous
160 solution, and the non-retained sulfonamides were determined by HPLC-UV analysis. For
161 kinetics studies, a 5 mg/L (each) sulfonamide aqueous solution was mixed with the C/Co-
162 SIM-1 carbon (1 mg/mL) under stirring at pH 5 and the concentration of remaining
163 sulfonamides in solution was measured every 30 minutes. The sorption data were analyzed
164 with a pseudo-second-order adsorption model,²⁵ whose linearized-integral form is expressed
165 by the following equation:

$$166 \quad \frac{t}{q_t} = \frac{t}{q_e} + \frac{1}{k_2 q_e^2}$$

167 where q_t and q_e (mg/g) are the amount of sulfonamide adsorbed at a time t (min) and at
168 equilibrium, respectively, and k_2 is the pseudo-second-order rate constant (g/mg min).

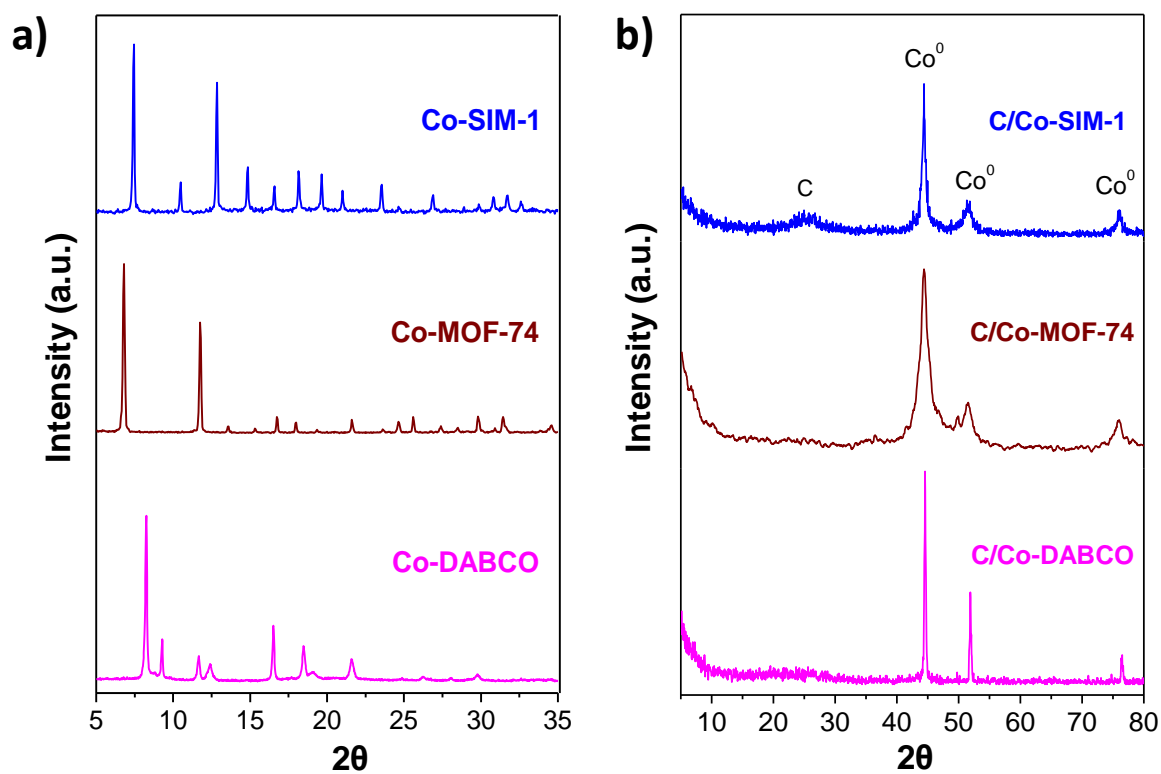
169 The enrichment factors were determined as the ratio of the peak areas from the HPLC
170 chromatograms before and after the extraction of the sulfonamides using 1 mL of methanol
171 as eluent. The reusability of the material was studied following the procedure described
172 before five times. Between consecutive extractions, the material was washed thoroughly with
173 methanol before being reused in the adsorption process. All the adsorption experiments were
174 carried out in triplicate.

175 **3. Result and Discussions**

176 **3.1 Preparation of magnetic porous carbons**

177 In a first step, Co-SIM-1, Co-MOF-74 and Co-DABCO MOFs precursors were synthesized
178 by solvothermal method. The X-ray diffractograms of all the cobalt metal-organic
179 frameworks (Fig. 1a) showed good crystallinity and matched well with the theoretical
180 diffraction patterns previously reported for the same MOF structures,^{22,24,26} demonstrating
181 that, in all cases, the corresponding MOF structure was obtained. Fig. 1b shows the
182 diffractograms of C/Co-SIM-1, C/Co-MOF-74 and C/Co-DABCO carbon samples. After the
183 carbonization process, no peaks corresponding to the precursor structure were detected in all
184 cases, while three new peaks at $2\theta = 43^\circ$, 53° and 76° appear in all the diffraction patterns,
185 which were assigned to metallic cobalt particles,^{27,28} although the presence of a small fraction
186 of amorphous cobalt oxide cannot be discarded. It should be noted that the width of the peaks
187 assigned to metallic cobalt particles differs from one sample to another, suggesting that the
188 MOF composition affects the size of the cobalt particles formed during the calcination
189 process. In addition, in the case of C/Co-SIM-1, a weak diffraction line at $2\theta = 25^\circ$ was
190 observed, which was assigned to graphitic carbon.^{28,29} The degradation of the prepared Co
191 containing MOFs was studied by thermogravimetric analysis (Figure S1). In all cases, the
192 first weight loss before 200°C is due to the removal of adsorbed solvents, and the second
193 weight loss around 300°C for Co-DABCO and 400°C for Co-SIM-1 and Co-MOF-74 is
194 attributed to the decomposition of MOFs structure, indicating the high thermal stability of
195 the precursors. After 800°C , the mass of the materials was relatively stable. The morphology
196 of the obtained carbons was studied by scanning electron microscopy. As it can be observed
197 in the corresponding SEM micrographs (Fig. 2a-c), the C/Co-SIM-1, C/Co-MOF-74 and
198 C/Co-DABCO samples are formed by micrometer-sized particles with cubic, hexagonal rod
199 and square rod morphology, respectively. Compared to the initial MOFs (Fig. S2), all the
200 derived carbons retained the original shape after the carbonization procedure due to the use

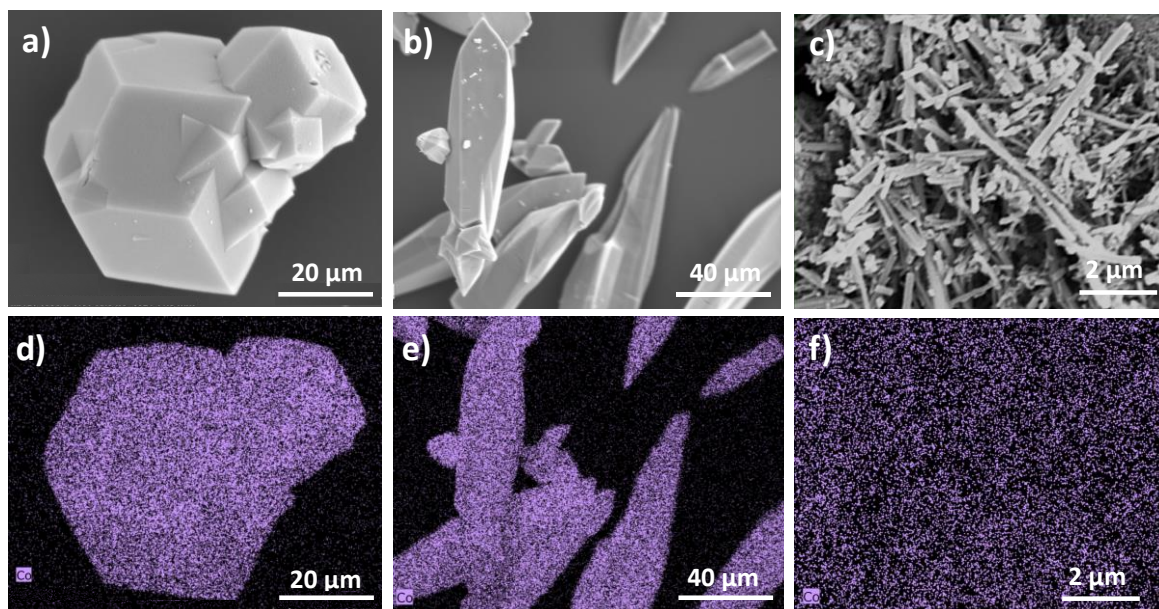
201 of thermally stable frameworks as precursors.^{30,31} Furthermore, elemental EDS mappings
202 (Fig. 2d-e) confirmed that, in all cases, the Co particles were homogeneously distributed on
203 the surface of the obtained carbon.



204

205 **Fig. 1.** X-ray diffractograms of (a) Co-SIM-1, Co-MOF-74 and Co-DABCO MOF samples
206 and (b) C/Co-SIM-1, C/Co-MOF-74 and C/Co-DABCO carbon samples.

207

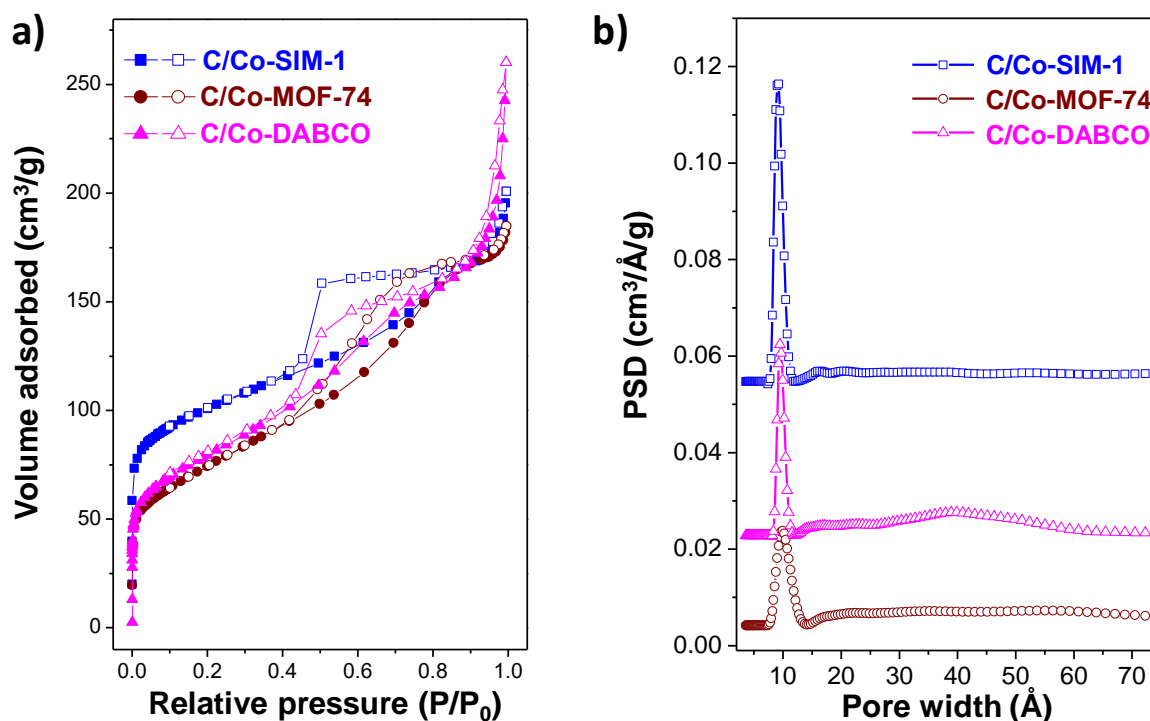


209

210 **Fig. 2.** SEM images and Co EDS mapping of (a,d) C/Co-SIM-1, (b,e) C/Co-MOF-74, and
 211 (c,f) C/Co-DABCO samples.

212

213 Nitrogen adsorption-desorption isotherms, and the corresponding pore size distribution
 214 curves, of the prepared carbon samples are shown in Fig. 3. All carbonaceous materials
 215 showed a significant nitrogen uptake at lower P/P_0 values (<0.1) and a hysteresis loop (Fig.
 216 3a), which indicate the simultaneous presence of micropores and mesopores.³² The
 217 determined BET surface areas ranged from 261 to 361 m^2/g (Table 1), which are lower than
 218 those of precursor MOFs (between 1213 and 373 m^2/g) due to a partial collapse of the
 219 structure during the carbonization process. The pore size distributions calculated using the
 220 2D-NLDFT model (Fig. 3c), corroborated, in all the cases, the coexistence of micropores,
 221 with a diameter centered around 10 Å, and mesopores with a broad size distribution (Fig. 3b
 222 and Table 1).



223

224 **Fig. 3.** (a) N₂ adsorption-desorption isotherms and (b) pore size distributions of C/Co-SIM-
 225 1, C/Co-MOF-74 and C/Co-DABCO samples.

226

227 **Table 1.** Textural properties of C/Co-SIM-1, C/Co-MOF-74 and C/Co-DABCO samples.

Sample	S _{BET} (m ² /g)	V _p (cm ³ /g)	Pore width (Å)
C/Co-SIM-1	361	0.25	8-10/14-100
C/Co-MOF-74	261	0.24	8-13/15-90
C/Co-DABCO	291	0.25	9-11/14-65

228

229

230 To study the nature of the carbon formed during the carbonization process, Raman spectra of
 231 the samples were recorded. As shown in Fig. 4a, all the carbon samples exhibited two broad
 232 bands near 1310 and 1581 cm⁻¹ which correspond to the D and G bands of the carbon,
 233 respectively. D band is attributed to the presence of defects or disorder within the carbon,
 234 and the G band is associated with the sp² hybridization existing in graphitic carbon.³³ The
 235 relative intensity ratio between D and G band (I_D/I_G) in the Raman spectra provides useful

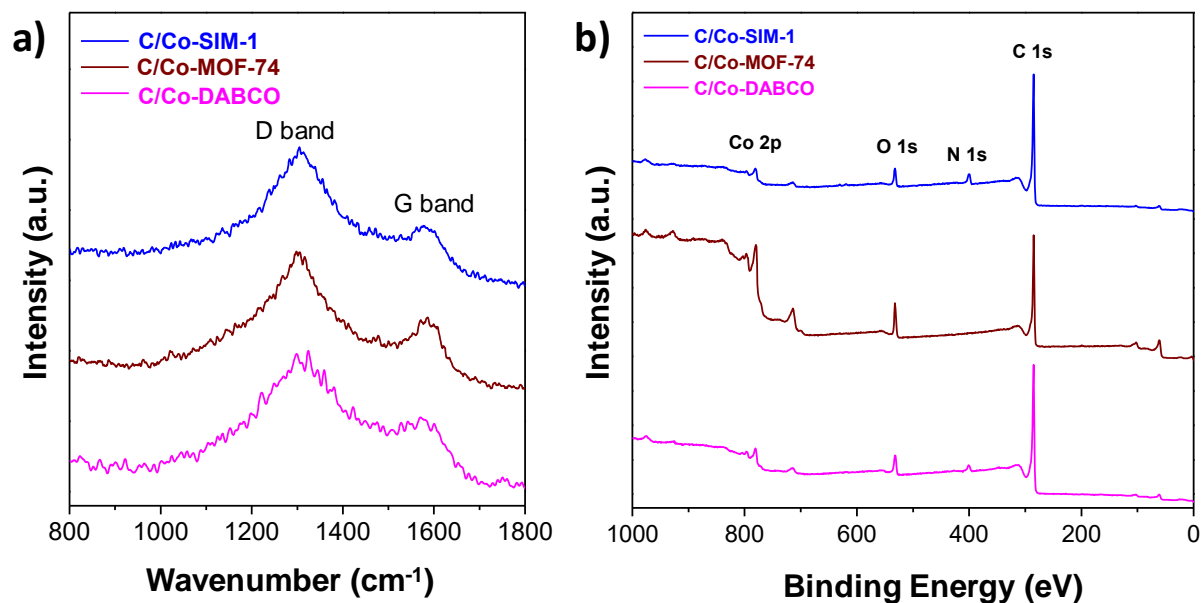
236 information about the graphitization degree of the carbonaceous materials.³⁴ The I_D/I_G ratios
237 calculated for C/Co-SIM-1, C/Co-MOF-74 and C/Co-DABCO were 2.21, 1.78 and 1.55,
238 respectively, indicating that the prepared carbons possessed mainly disorder structures or
239 contained long-range disordered graphitic carbon, being the C/Co-SIM-1 the carbon which
240 presented more quantity of defects. X-ray photoelectron spectroscopy (XPS) analysis of the
241 MOFs derived carbons was performed to investigate their chemical composition. In all cases,
242 the spectra (Fig. 4b) exhibit peaks in the C 1s, Co 2p and O 1s regions, which confirm the
243 presence of these elements in all the materials. Additionally, the spectra of C/Co-SIM-1 and
244 C/Co-DABCO showed a signal in the N 1s region, indicating the presence of N in both
245 samples coming from the ligands used for their synthesis. For all the materials the high-
246 resolution C 1s spectra (Fig. S3) showed two peaks centered at 284.5 eV and 288.5 eV, which
247 are assigned to C–C/C=C and C=O bonds, respectively.^{35,36} The spectra in the Co 2p region
248 of C/Co-SIM-1, C/Co-MOF-74 and C/Co-DABCO (Fig. S4) are fitted to six peaks
249 corresponding to metallic cobalt (794.8 eV - 793.8 eV and 779.6 eV - 778.6 eV), cobalt oxide
250 (796.8 eV – 796.4 eV and 781.2 eV – 780.6 eV) and shake-up satellite peaks (803.8 eV –
251 803.0 eV and 785.8 eV – 785.0 eV).^{35,37,38} The spectra N 1s of C/Co-SIM-1 and C/Co-
252 DABCO (Fig. S5) showed three signals centered at 398.6, 400.7 and 405.4 eV, which can be
253 assigned to pyridinic N, graphitic N and oxidized N, respectively.³⁹

254

255

256

257



258 **Figure 4.** (a) Raman and (b) XPS spectra of C/Co-SIM-1, C/Co-MOF-74 and C/Co-DABCO
 259 samples.

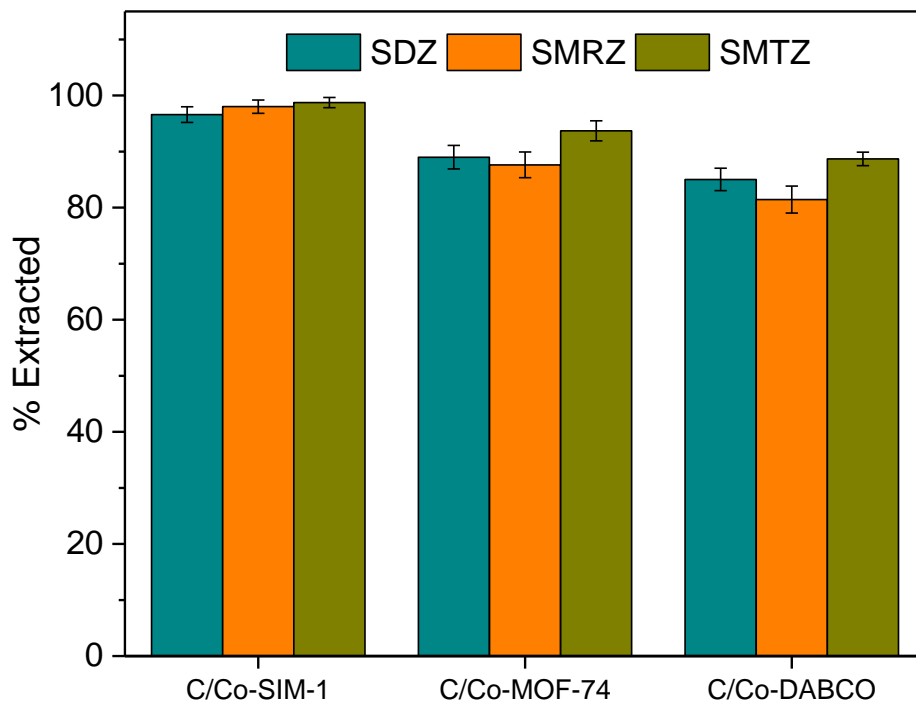
260

261 3.2 Extraction of sulfonamides

262

263 Carbons derived from cobalt (II)-based metal-organic frameworks were tested for the
 264 simultaneous magnetic extraction of three sulfonamides (sulfadiazine, sulfamerazine and
 265 sulfametazine). Fig. 5 shows the percentages of sulfonamides adsorbed by the three prepared
 266 carbons from a 5 mg L⁻¹ (each) solution. As it can be observed, the three MOF-derived
 267 carbons showed high extraction percentages, which are comparable or even better than most
 268 of the values reported in the literature using other magnetic adsorbents,^{40–42} especially in the
 269 case of C/Co-SIM-1 carbon which reaches values higher than 95%. The excellent adsorption
 270 performance of this material could be attributed to its higher surface area and bimodal pore
 271 structure, which facilitates the access of sulfonamides to the pores, and its significant amount
 272 of surface defects, which could act as adsorption sites of SNs, improving the extraction

273 capacity.³⁶ According to these results, the C/Co-SIM-1 was selected for the study of the
274 extraction variables.



275

276 **Fig. 5.** Percentage of sulfonamides extracted by C/Co-SIM-1, C/Co-MOF-74 and C/Co-
277 DABCO samples.

278

279 The leaching of cobalt ions into the water after the adsorption experiments using C/Co-SIM-
280 1 was studied by inductively coupled plasma optical emission spectrometry (ICP-OES). After
281 2 h in solution, a small amount of cobalt was detected in the supernatant solution, which
282 remained constant with the time, indicating that after the carbonization treatment of the MOF
283 a small quantity of cobalt is not well attached to the carbon matrix. To eliminate it, the carbon
284 sample was washed with water before the extraction experiments.

285 Taking into account that the pH of the solution influences adsorption, since it determines the
286 surface charge of adsorbent and the speciation of the adsorbate, the extraction of

287 sulfonamides by the C/Co-SIM-1 carbon was carried out at different pH values (Fig. S6). As
288 it can be observed, the highest extraction percentages for the three SNs were achieved at pH
289 values between 5 and 7.5 and the adsorption capacity significantly decreased when the pH
290 increased from 7.5 to 10. Considering the pKa values of the studied sulfonamides⁴ and the
291 isoelectric point of C/Co-SIM-1 (Fig. S7), it must exist an electrostatic attraction between
292 the negatively charged surface of the carbon material and the positively charged SNs at weak
293 acid and neutral pH, which becomes a repulsion in basic conditions due to the negative charge
294 of SNs at high pH values. However, the high extraction capacity of C/Co-SIM-1, even at
295 basic pH, suggests that other mechanisms, such as π - π interactions and hydrogen bonding,
296 should exist between the carbon and the sulfonamides.^{10,43}

297 The effect of the contact time on the adsorption of SDZ, SMRZ and SMTZ on C/Co-SIM-1
298 was also studied (Fig. S8). For the three sulfonamides, the adsorption equilibrium was
299 reached in 30 min, indicating a fast extraction process. The adsorption kinetic data were
300 analyzed using a pseudo-second-order kinetic model;²⁵ the fitting results are presented in
301 Table 2. The obtained determination coefficients (R^2), which were between 0.9916 and
302 0.9932, indicate that the experimental data could be successfully fitted by this model,
303 suggesting that chemical adsorption is the rate-limiting step of the adsorption process and it
304 is dependent on the number of active sites.⁸ The amount of each of the sulfonamides adsorbed
305 at equilibrium ranks among the highest ones reported in the literature using magnetic
306 adsorbents, demonstrating the high adsorption capacity of C/Co-SIM-1,⁴⁴⁻⁴⁹ probably related
307 to its high surface area and porosity and the different interactions that could take place
308 between the adsorbent and the target analytes.

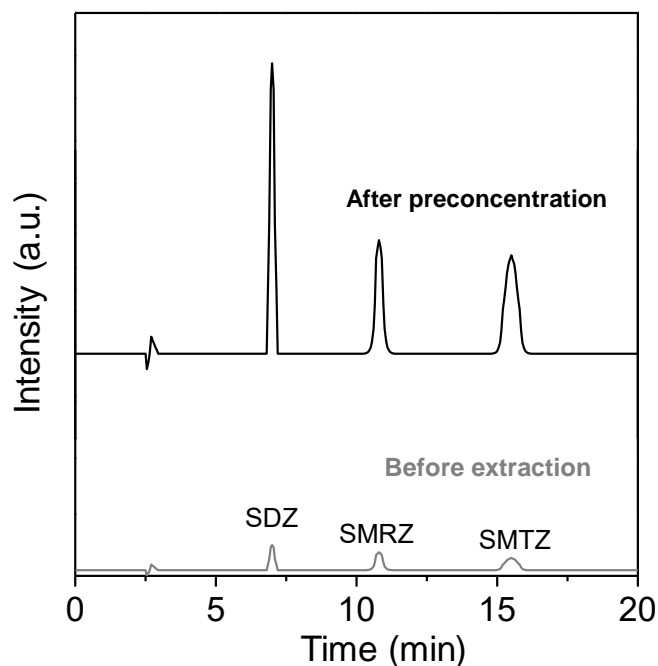
309

310 **Table 2.** Parameters of the pseudo-second-order kinetic model for the SNs adsorption by
311 C/Co-SIM-1 carbon.

SNs	$q_{e(\text{exp})}$ (mg/g)	Pseudo-second-order kinetic model		
		$q_{e(\text{calc})}$ (mg/g)	k_2 (g/mg min)	R^2
SDZ	48.21	44.44	0.00111	0.9916
SMRZ	48.85	45.04	0.00113	0.9932
SMTZ	49.51	45.66	0.00112	0.9931

312

313 In order to study the applicability of the C/Co-SIM-1 carbon, we evaluated its capacity for
314 the simultaneous extraction and preconcentration of sulfonamides from water. Fig. 6 shows
315 the chromatograms obtained, before and after solid-phase extraction, of the three
316 sulfonamides from a mixture solution of 5 mg/L each, using C/Co-SIM-1 as adsorbent and 1
317 mL of methanol as eluent. It can be seen that the intensity of the SNs peaks significantly
318 increases after preconcentration, reaching enrichment factors of 47, 43 and 44 for SDZ,
319 SMRZ and SMTZ, respectively, in comparison with the direct injection of a mixture of
320 sulfonamides.



321

322 **Fig. 6.** SDZ, SMRZ and SMTZ chromatograms before and after solid-phase extraction by
 323 using C/Co-SIM-1.

324

325

326 An important feature for practical extraction applications is the reusability of the adsorbent.

327 To evaluate the capacity of regeneration of the C/Co-SIM-1 material, a recyclability

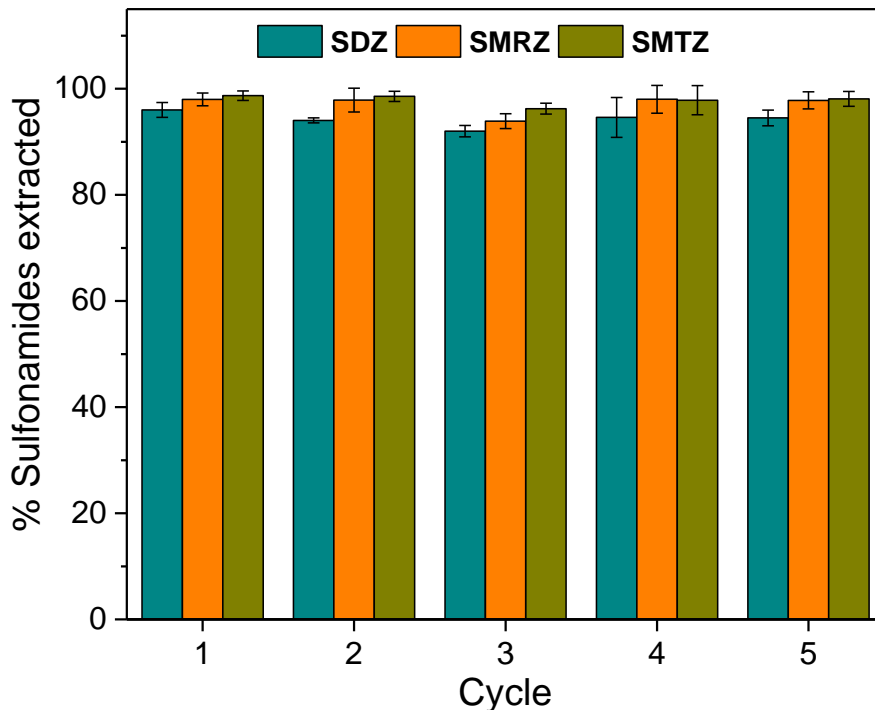
328 extraction test during five consecutive cycles was carried out. As it can be observed in Fig.

329 7, after 5 extraction cycles, similar extraction capacities of the three sulfonamides were

330 obtained. The relative standard deviations of the extraction capacity of SDZ, SMRZ and

331 SMTZ were 1.5, 2.1 and 1.0%, respectively, demonstrating the excellent reusability of C/Co-

332 SIM-1 for the extraction of sulfonamides.



333

334 **Figure 7.** Recyclability of C/Co-SIM-1 for adsorption of SNs from water.

335

336 **4. Conclusions**

337

338 In this study, the use of magnetic porous carbons derived from metal-organic frameworks as
 339 sorbents for solid-phase extraction of sulfonamides has been explored for the first time. Three
 340 different cobalt(II)-based MOFs: Co-SIM-1, Co-MOF-74 and Co-DABCO, were used as
 341 precursors for the preparation of highly porous magnetic carbons via a simple carbonization
 342 process. These magnetic carbons were used for the dispersive solid-phase extraction of
 343 sulfamethazine, sulfamerazine and sulfadiazine avoiding tedious filtration or centrifugation
 344 steps. The highest SNs extraction performance was reached with C/Co-SIM-1 carbon, due to
 345 its large number of surface defects, high surface area and bimodal pore structure which
 346 facilitated the SNs adsorption on the internal channels of this material. The C/Co-SIM-1

347 exhibited fast extraction kinetics and high enrichment factors ranging from 43 to 47 for the
348 mixture of three sulfonamides catalogued as emerging contaminants. The obtained C/Co-
349 SIM-1 carbon was reused several times keeping its extraction performance, making it a
350 promising material for the simultaneous extraction and preconcentration of emerging
351 pollutants from water.

352

353 **5. Acknowledgements**

354 The authors acknowledge Spanish Agencia Estatal de Investigación (AEI) and
355 European Funds for Regional Development (FEDER) for financial support through
356 Project CTQ2016-77155-R (AEI/FEDER, UE). Mendiola-Alvarez thanks to
357 CONACYT and to Centro de Estudios de Posgrado (UIB) for the scholarship support.
358 The authors acknowledge F. Hierro Riu for scanning micrographs.

359

360 **6. References**

- 361 1. Y. Liu, J. Wang, *J. Hazard Mater.*, 2013, **250–251**, 99–105.
- 362 2. M. García, M. Villagrasa, M. Díaz, D. Barceló, *Anal. Bioanal. Chem.*, 2010, **397**,
363 1325–1334.
- 364 3. K. Ikehata, N. Jodeiri, M. Gamal, *Ozone Sci. Eng.*, 2006, **28**, 353–414.
- 365 4. Y. Ma, K. Zhang, C. Li, T. Zhang, N. Gao, *Biomed. Res. Int.*, 2015, **2015**, 1–10.
- 366 5. A. Batt, D. Snow, D. Aga, *Chemosphere*, 2006, **64**, 1963–1971.
- 367 6. Q. Jiuhui, *J. Environ. Sci.*, 2008, **20**, 1–13.
- 368 7. A. Mansoor, M. Lashgari, *Chem. Eng. J.*, 2009, **150**, 555–560.
- 369 8. Y. Gao, R. Kang, J. Xia, G. Yu, S. Deng, *J. Colloid Interface Sci.*, 2019, **535**, 159–
370 168.
- 371 9. M. Azhar, H. Abid, V. Periasamy, H. Sun, M. Tade, S. Wang, *J. Colloid Interface*
372 *Sci.*, 2017, **500**, 88–95.

- 373 10. S. Hasan, A. Easley, M. Beth, B. Monroe, D. Maitland, *J. Colloid Interface Sci.*, 2016,
374 **478**, 334–343.
- 375 11. Y. Shih, K. Wang, B. Singco, C. Lin, H. Huang, *Langmuir*, 2016, **32**, 11465–11473.
- 376 12. B. Hashemi, P. Zohrabi, N. Raza, K. Kim, *TrAC - Trends Anal. Chem.*, 2017, **97**, 65–
377 82.
- 378 13. S. Yang, T. Kim, J. Im, Y. Kim, K. Lee, H. Jung, C. Park, *Chem. Mater.*, 2012, **24**,
379 464–470.
- 380 14. L. Yang, X. Zeng, W. Wang, D. Cao, *Adv. Funct. Mater.*, 2018, **28**, 1704537.
- 381 15. C. Palomino Cabello, M. Picó, F. Maya, M. del Rio, G. Turnes Palomino, *Chem. Eng.*
382 *J.*, 2018, **346**, 85–93.
- 383 16. B. Bhadra, S. Jhung, *Microporous Mesoporous Mater.*, 2018, **270**, 102–108.
- 384 17. A. Banerjee, R. Gokhale, S. Bhatnagar, J. Jog, M. Bhardwaj, B. Lefez, *J. Mater.*
385 *Chem.*, 2012, **22**, 19694–19699.
- 386 18. L. Wang, F. Ke, J. Zhu, *Dalt. Trans.*, 2016, **45**, 4541–4547.
- 387 19. Y. Liu, Z. Gao, R. Wu, Z. Wang, X. Chen, T. Chan, *J. Chromatogr. A*, 2017, **1479**,
388 55–61.
- 389 20. F. Maya, C. Palomino Cabello, R. Frizzarin, J. Estela, G. Turnes Palomino, V. Cerdà,
390 *TrAC - Trends Anal. Chem.*, 2017, **90**, 142–152.
- 391 21. S. Zhang, Q. Yang, Z. Li, W. Wang, C. Wang, Z. Wang, *Analyst*, 2016, **141**, 1127–
392 1135.
- 393 22. S. Aguado, J. Quirós, J. Canivet, D. Farrusseng, K. Boltes, R. Rosal, *Chemosphere*,
394 2014, **113**, 188–192.
- 395 23. H. Cho, D. Yang, J. Kim, S. Jeong, W. Ahn, *Catal. Today*, 2012, **185**, 35–40.
- 396 24. L. Gómez, S. Castro, M. Sánchez, S. Yáñez-, J. Mira, J. Bermúdez, T. Centeno, M.
397 Señarís, *Eur. J. Inorg. Chem.*, 2016, **2016**, 4463–4469.
- 398 25. Y. Ho, G. McKay, *Process Biochem.*, 1999, **34**, 451–465.
- 399 26. N. Rosi, J. Kim, M. Eddaoudi, B. Chen, M. O’Keeffe, O. Yaghi, *J. Am. Chem. Soc.*,
400 2005, **127**, 1504–1518.
- 401 27. X. Li, C. Zeng, J. Jiang, L. Ai, *J. Mater. Chem. A*, 2016, **4**, 7476–7482.
- 402 28. N. Torad, M. Hu, S. Ishihara, H. Sukegawa, A. Belik, M. Imura, K. Ariga, Y. Sakka,
403 Y. Yamauchi, *Small*, 2014, **10**, 2096–2107.
- 404 29. S. Chen, L. Wen, F. Svec, T. Tan, Y. Lv, *RSC Adv.*, 2017, **7**, 21205–21213.

- 405 30. Z. Yu, X. Wang, Y.-N. Hou, X. Pan, Z. Zhao, J. Qiu, *Carbon*, 2017, **117**, 376–382.
- 406 31. S. Dang, Q. Zhu, Q. Xu, *Nat. Rev. Mater.*, 2018, **3**, 17075.
- 407 32. M. Thommes, K. Kaneko, A. Neimark, J. Olivier, F. Rodriguez, J. Rouquerol, K.
408 Sing, *Pure Appl. Chem.*, 2015, **87**, 1051–1069.
- 409 33. C. Zhang, N. Mahmood, H. Yin, F. Liu, Y. Hou, *Adv. Mater.*, 2013, **25**, 4932–4937.
- 410 34. B. Chen, G. Ma, Y. Zhu, Y. Xia, *Sci. Rep.*, 2017, **7**, 1–9.
- 411 35. C. Cui, Z. Wei, G. Zhou, W. Wei, J. Ma, L. Chen, C. Li, *J. Mater. Chem. A*, 2018, **6**,
412 7088–7098.
- 413 36. Q. Wang, L. Zhang, *J. Chromatogr. A*, 2018, **1568**, 1–7.
- 414 37. S. Das, M. Bhunia, M. Motin, S. Dutta, A. Bhaumik, *Dalt. Trans.*, 2011, **40**, 2932–
415 2939.
- 416 38. H. Lu, H. Zhang, R. Liu, X. Zhang, H. Zhao, G. Wang, *Appl. Surf. Sci.*, 2017, **392**,
417 402–409.
- 418 39. L. Wang, S. Dou, J. Xu, H. Liu, S. Wang, J. Ma, S. Dou, *Chem. Commun.*, 2015, **51**,
419 11791–11794.
- 420 40. H. Wu, Y. Shi, X. Guo, S. Zhao, J. Du, H. Jia, L. He, L. Du, *J. Sep. Sci.*, 2016, **39**,
421 4398–4407.
- 422 41. p-S. Gao, Y. Guoa, X. Lia, X. Wang, J. Wang, F. Qian, H. Gu, Z. Zhang, *J.*
423 *Chromatogr. A*, 2018, **1575**, 1–10.
- 424 42. Y. Zhao, R. Wu, H. Yu, J. Li, L. Liu, S. Wang, X. Chen, T.-W. D. Chan, *J.*
425 *Chromatogr. A*, 2020, **1610**, 460543.
- 426 43. M. Teixidó, J. Pignatello, J. Beltrán, M. Granados, J. Peccia, *Environ. Sci. Technol.*,
427 2011, **45**, 10020–10027.
- 428 44. P. Shi, N. Ye, *Anal. Methods*, 2014, **6**, 9725–9730.
- 429 45. M. Izanloo, A. Esrafil, M. Behbahani, M. Ghambarian, H. Reza Sobhi, *J. Sep. Sci.*,
430 2018, **41**, 910–917.
- 431 46. Y. Li, N. Zhu, T. Chen, Y. Ma, Q. Li, *Microchem. J.*, 2018, **138**, 401–407.
- 432 47. J. Zhang, Z. Chen, S. Tang, X. Luo, J. Xi, Z. He, J. Yu, F. Wu, *Anal. Chim. Acta*,
433 2019, **1089**, 66–77.
- 434 48. J.-M. Liu, S.-W. Lv, X.-Y. Yuan, H.-L. Liu, S. Wang, *RSC Adv.*, 2019, **9**, 14247–
435 14253.
- 436 49. Z. Chen, Z. He, X. Luo, F. Wu, S. Tang, J. Zhang, *Food Anal. Method*, 2020,
437 <https://doi.org/10.1007/s12161-020-01750-2>.

438

439

440

441

442

443

444

445

446

447

448

449

450

451

452

453